



Review

Solid waste management in European countries: A review of systems analysis techniques

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ABSTRACT

In the past few decades, solid waste management systems in Europe have involved complex and multi-faceted trade-offs among a plethora of technological alternatives, economic instruments, and regulatory frameworks. These changes resulted in various environmental, economic, social, and regulatory impacts in waste management practices which not only complicate regional policy analysis, but also reshape the paradigm of global sustainable development. Systems analysis, a discipline that harmonizes these integrated solid waste management strategies, has been uniquely providing interdisciplinary support for decision making in this area. Systems engineering models and system assessment tools, both of which enrich the analytical framework of waste management, were designed specifically to handle particular types of problems. Though how to smooth out the barriers toward achieving appropriate systems synthesis and integration of these models and tools to aid in the solid waste management schemes prevalent in European countries still remains somewhat uncertain. This paper conducts a thorough literature review of models and tools illuminating possible overlapped boundaries in waste management practices in European countries and encompassing the pros and cons of waste management practices in each member state of the European Union. Whereas the Southern European Union (EU) countries need to develop further measures to implement more integrated solid waste management and reach EU directives, the Central EU countries need models and tools with which to rationalize their technological choices and management strategies. Nevertheless, considering systems analysis models and tools in a synergistic way would certainly provide opportunities to develop better solid waste management strategies leading to conformity with current standards and foster future perspectives for both the waste management industry and government agencies in European Union.

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Abbreviations: ADEME, Agence de l'Environnement et de la Maîtrise de l'Energie; AWAST, aid in the management and European comparison of a municipal solid waste treatment for a global and sustainable approach; BMW, biodegradable municipal waste; CBA, cost–benefit analysis; CERA, comparative environmental risk assessment; DP, dynamic programming; DSD, Duales System Deutschland; DSSs, decision support systems; EASEWASTE, environmental assessment of solid waste systems and technologies; EC, European Community; EDX/EDI, electronic data exchange; EEA, European Environment Agency; EEC, European Economic Community; EIA, environmental impact assessment; EIONET, European environment information and observation network; EPR, extended producer responsibility; ERA, environmental risk assessment; ES, expert system; EU, European Union; EUDIN, European Data Interchange of Waste Notification System; FM, forecasting models; GHG, greenhouse gas; GIGO, garbage in, garbage out; GIP, grey integer programming; GIS, geographic information system; IMS, integrated modeling system; IOA, input–output analysis; ISWM, integrated solid waste management; IWM, integrated waste management; LATS, landfill allowance trading system; LCA, life cycle assessment; LCI, life cycle inventory; LP, linear programming; MCDM, multicriteria decision making; MFA, material flow analysis; MIP, mixed-integer programming; MIMES/WASTE, model for description and optimization of integrated material flows and energy systems; MIS, management information system; MSW, municipal solid waste; NIMBY, not in my backyard; NLP, non-linear programming model; OM, optimization models; ORWARE, ORganic WASTE REsearch; PAYT, pay-as-you-throw; PET, polyethylene terephthalate; QAS, quality assurance system; RA, environmental and ecological risk assessment; SA, sustainability assessment; SD, scenario development; SDS, sustainable development strategy; SEA, strategic environmental assessment; SFA, substance flow analysis; SFINX, substance flow inter-nodal exchange; SM, simulation models; SoEA, socioeconomic assessment; STAN, subStance flow ANalysis; SWIM, solid waste-integrated model; SWM, solid waste management; TASAR, tool for analyzing separation actions and recovery; WASTED, waste analysis software tool for environmental decisions; WHP, waste hierarchy principle; WISARD, waste-integrated systems for assessment of recovery and disposal; WRAP, Waste Resources Allocation Program; WRATE, waste and resources assessment tool for the environment; XML, extensible markup language.

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1. Introduction

In the 21st century, the sustainable management of municipal solid waste (MSW) will become necessary at all phases of impact from planning to design, to operation, and to decommissioning. As a consequence, the spectrum of new and existing waste treatment technologies and managerial strategies has also spanned from maintaining environmental quality at present to meet sustainability goals in the future. Such an orderly evolution allows both waste management industries and government agencies to meet common needs of waste management with greatest green potential, to recycle materials out of waste streams, to enlarge the renewable energy supply, to seek for more socially acceptable options, and to preserve biodiversity and natural ecosystems simultaneously. To achieve such goals, all technical and non-technical aspects of a solid waste management (SWM) system should be analyzed as a whole, since they are inter-related with one another and developments in one area frequently affect practices or activities in another area (UNEP, 2005).

Systems analysis techniques have been applied to handle MSW streams through a range of integrative methodologies in the last few decades. A total of five system engineering models and nine system assessment tools were formally classified in this field to illuminate the challenges, trends and perspectives (Chang et al., *in press*). It is worth knowing that the spectrum of these models and assessment tools was classified based on the following two domains although some of them may be intertwined with each other (Chang et al., *in press*). They are: 1) systems engineering models including cost–benefit analysis (CBA), forecasting models (FM), simulation models (SM), optimization models (OM), and integrated modeling system (IMS), as well as 2) system assessment tools including management information system (MIS)/decision support system (DSS)/expert system (ES), scenario development (SD), material flow analysis (MFA), life cycle assessment or life cycle inventory (LCA or LCI), risk assessment (RA), environmental impact assessment (EIA), strategic environmental assessment (SEA), socioeconomic assessment (SoEA), and sustainable assessment (SA). Fig. 1 holistically illustrates the interrelationships among these two domains from which fourteen technologies can be connected through such a technology hub in association with these two broad-based domains (Chang et al., *in press*). In the core part, the five systems engineering models can be seen as the core technologies in which the cost–benefit analysis may be used as a common platform in support of decision making. Integrated modeling systems may flexibly concatenate various optimization models including linear programming (LP), mixed-integer programming (MIP), non-linear programming (NLP), and dynamic programming (DP) models to address the system concerns in which the SM and FM can support the essential background in concert with CBA in the context of systems analysis. With such a core structure, the model-based DSSs can be constructed for separate or collective applications. Yet rule-based, knowledge-based or graphics-based DSSs or ESs can still be formed based on heuristic approaches. All of these core efforts may be enhanced by the rest of system assessment tools described by the eight outer triangles. Communication among the eight triangles canalizes the information flows that in turn improve the credibility of the five systems engineering models being formulated through MIS, DSS, and even ES. Overall, Fig. 1 leads to a sound realization of the structure between systems engineering models and systems assessment tools from which a systems analysis should be well balanced for generating environmentally benign, cost effective, ecologically sound, and socially acceptable solutions (Morrissey and Browne, 2004; Chang and Davila, 2007).

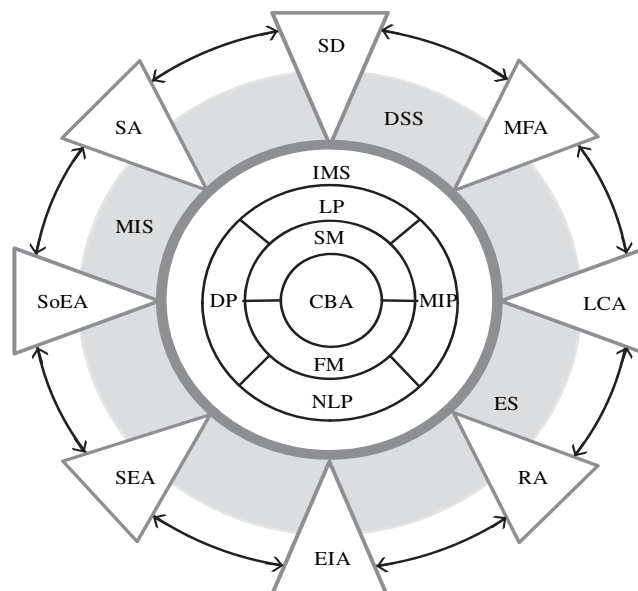


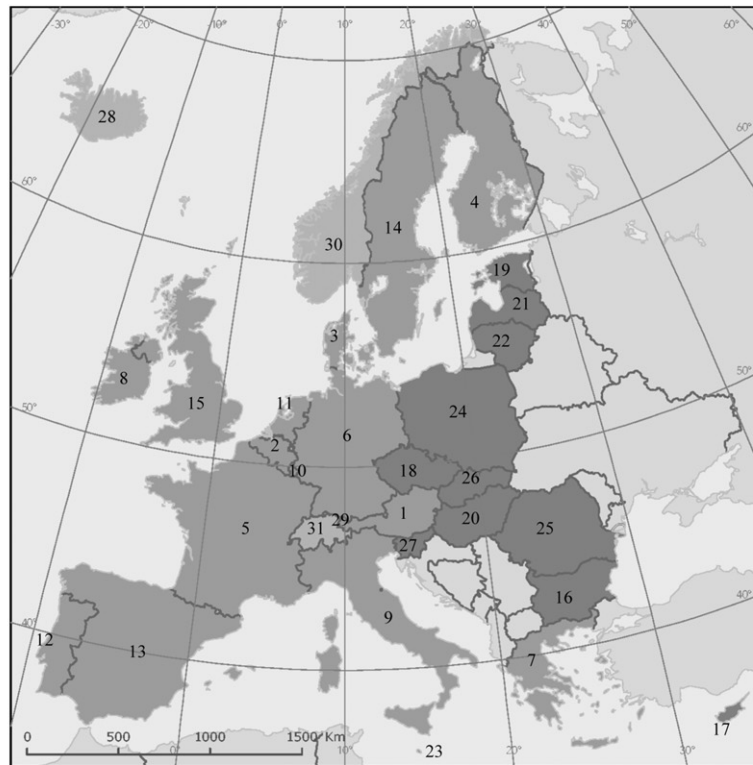
Fig. 1. The technology hub for solid waste management (Chang et al., 2009).

With such a tool, every community can tailor its own unique system to manage various components of the waste streams in a flexible manner (Najm et al., 2002). Yet how to smooth out the barriers toward achieving appropriate systems synthesis and integration of the five systems engineering models and the nine system assessment tools to aid in solid waste management practices in European countries remains somewhat uncertain. It is the aim of this paper to present a thorough literature review and a critical analysis in sequence so as to answer the following key questions: 1) what achievements have been reached so far?, 2) what are the gaps in knowledge of waste management that we need to achieve in the context of sustainable development in the long run? and 3) what are the research needs and future directions in systems analysis for SWM in European countries. At a practical level, discussions of this paper were limited to 15 European Union (EU) member states, facing the same driving forces with similar waste legislation to manage MSW systems. The EU is an economic and political union of 27 member states which are located primarily in Europe, Norway and Switzerland (non-EU members) were also included to understand how other countries within the region with similar waste management legislation applied the techniques of systems analysis to manage their waste management issues. Fig. 2 illustrates the overall study boundaries.

The comparative analysis of this paper provides an all-inclusive view to minimize anomalies of SWM systematically, and should allow possible conflicts associated with different objectives associated with environmental, social, technical, and economic constraints to be confronted more rationally than heretofore. Such a development should enable more solidly based waste management strategies to be pursued, leading to conformity with current standards for both the waste management industry and government agencies in the EU.

2. Current waste management principles in the EU

After the commitments made at the Earth Summit in Rio de Janeiro (1992), the European Council in 2001 adopted the first EU Sustainable Development Strategy (SDS). The overall aim of the renewed EU SDS is to support and promote actions enabling the EU to achieve



EU-15		New Member States		European Free Trade Association	Other European Countries
1. Austria	9. Italy	16. Bulgaria	22. Lithuania	28. Iceland	
2. Belgium	10. Luxembourg	17. Cyprus	23. Malta	29. Liechtenstein	
3. Denmark	11. Netherlands	18. Czech Republic	24. Poland	30. Norway	
4. Finland	12. Portugal	19. Estonia	25. Romania	31. Switzerland	
5. France	13. Spain	20. Hungary	26. Slovakia		
6. Germany	14. Sweden	21. Latvia	27. Slovenia		
7. Greece	15. UK				
8. Ireland					

Fig. 2. Groups of countries within the European Union (adapted from EEA, 2007).

continuous improvement of quality of life for both current and future generations. This is expected to be achieved through the creation of sustainable communities capable of managing resources efficiently, tapping the innovation potential of the economy, ensuring prosperity, environmental protection and social cohesion. These changes will bring about a sense of urgency in SWM. While short-term action is required for tackling operational issues in SWM, maintaining a long-term perspective of SWM also needs to be set out. The most recent legislation published by European Commission (EC) is the New Waste Directive 2008/98/EC (EU, 2008), which reflects EU SDS and brings new challenges to SWM systems. New definitions for waste, by-products and end-of-waste, result in the need for choosing appropriate technologies that aim at improving the protection of human

health and environment, promoting reuse and recycling, enhancing waste prevention programs via biowaste separate collection, and implementing extended producer responsibility (EPR) collectively. In addition, key challenges related to long-term waste management are climate change and energy use, linking SWM systems with the reduction of greenhouse gas (GHG) emissions and the enhancement of energy recovery. Sustainable shipping of waste streams is thus an important issue in SWM too. Sustainable consumption and production, related to waste prevention programs have received wide attention in the nexus of resources conservation, recovery, and reuse. Social factors, including population growth and migration, become essential for the accurate forecasting of waste generation and estimation of the proper capacity of the SWM facilities. Public health,

which used to be considered by LCA impact categories must be included through the application of a quality assurance system (QAS) for product control. All of them compound the structure of current SWM systems and deepen the need for systems analysis within the EU member states.

Proper consideration of the impacts of climate changes and resources scarcity has been mandatory in environmental management including SWM in Europe. Scarcity of resources has motivated new strategies at European level to promote life cycle thinking in waste management policies, and consequently, the problems of MSW management are tied with how to integrate economically feasible and environmentally sustainable practices holistically. Challenges arise with respect to interfaces between optimal planning and sizing of solid waste management facilities and optimal scheduling of waste flows and throughputs while evaluating new system components and taking into consideration environmental and social costs, such as municipal taxes, user charges, capital opportunity costs and government grants and subsidies. These socioeconomic strategies, which were implemented only by a handful of industrialized countries in the world, might be extended to reduce waste generation and, simultaneously, de-link waste generation from economic growth.

An improved knowledge base influences the advancement of waste collection and shipping, resources use, and disposal alternatives via substitution of more systematic modeling practices. The Thematic Strategy on the Prevention and Recycling of Waste (EU, 2005) is an example of such a policy change. Improving the existing legislation, with simplification and modernization effects on waste definition, end-of-waste criteria, recycling, recovery and disposal activities, is one of the guidelines which is crucial to continue into the next decade. In addition, climate changes have also forced new measures to be implemented at the EU level. They include promoting GHG emissions reduction through biowaste diversion from landfills, improving energy efficiency at waste treatment and disposal facilities, promoting organic fertilizers (compost) in soils as an alternative to mineral fertilizers, enhancing quality in waste management outputs (like recycled materials) to reduce resource consumption, and raising materials' utility. Some social aspects in MSW management have also been made mandatory by EU regulations, like the SEA Directive, related to public participation with respect to the drawing up of certain plans and programs relevant to the environmental directive (EU, 2001).

To ultimately improve urban sustainability and offer the level of service required by the population, the ability to increase the reliability of green infrastructure systems with waste management functionalities, particularly through the proper interfaces between the partnerships of private and public sectors could be even more critical. Ultimately, not only mitigatory solutions related to climate change should comprise a part of the MSW management strategies but also challenges in adaptation are also made necessary for SWM, which are mainly related to waste treatment technologies. On the other hand, waste collection systems should be designed and operated so as to be capable of improving public health protection. For example, higher temperatures after climate change that may result in more biowaste degradation, thereby generating odor control problems which require further attention. More sophisticated societal measures, such as voluntary agreements on encouraging responsibility among producers and consumers, which might become mandatory to reach an integrated solid waste management (ISWM), compound the decision making when arriving at consensus and involvement between stakeholders and decision makers during participation processes in different EU member states. This situation triggers an acute need to thoroughly review all

the existing “state-of-the-art” systems engineering approaches and system assessment tools in order to reach such ISWM goals, particularly in EU member states with differing levels of economic development. These goals are necessary to facilitate a unique integration for tackling complexity with respect to multi-objectives, risk, and uncertainty characteristics in decision making as a whole.

3. Systems analysis techniques

In systems engineering regimes, a system can be a set of related components or sub-systems, which interact with each other in some way. The properties of a system are defined by the whole of the sub-systems, their characteristics, and their relationships. The characteristics are related to the boundaries of the system depending on whether they are closed or open systems/sub-systems. With this definition, an MSW management system fits the concept in which the technical aspects like landfill, incineration, anaerobic digestion, composting and collection are sub-systems linked with one another through processed waste streams internally and municipalities through managed truck fleets externally. The sub-systems make up part of the SWM system that has interactions between technical and non-technical aspects, both of which may influence the generation and shipping of waste to some extent.

Considering SWM systems, several systems analysis techniques have been applied to help decision making. These can be divided into two main groups as we mentioned above: systems engineering models and systems assessment tools. Their contribution to SWM systems will be summarized in the following sub-sections in concert with the concept.

3.1. Systems engineering models

Complexity in SWM system arises from siting facilities, selecting technologies, and comparing management options. To tackle the synergistic interfaces, systems engineering models can be helpful for promoting analysis based on cost–benefit analysis (CBA), optimization models (OM), simulation models (SM), forecasting models (SM) and integrated modeling systems (IMS). Table 1 presents the contribution of systems engineering models to SWM system analysis over the past few decades. It offers a systematic overview showing how the landscape of systems engineering models was conceptualized (complexes, structures, functionalities) and the relationship to other components in connection with Fig. 1.

3.2. Systems assessment tools

Most of the time, after systems have been created and implemented, it is necessary to evaluate their performance and consider how improvements could be made, especially in answer to the increasing challenges promoted by regulation. Models that can help decision makers toward such goals are systems assessment tools. Such tools can be MIS, DSS, ES, SD, MFA, LCA, RA, EIA, SEA, SoEA and SA. Table 2 presents the contribution of systems assessment tools to SWM with significant ramifications on how the relationships between system assessment tools and systems engineering models are described when connected with Fig. 1. As a consequence, the appropriate use of system assessment tools has such a large effect on the overall optimization especially in the context of IMS because the outputs from these tools are normally used as the primary inputs in models reflecting socioeconomic, climate change, and managerial considerations.

Table 1

The contribution of systems engineering models to SWM systems (Chang et al., in press).

Types of systems engineering models	Description	Contribution to SWM system
Cost–benefit analysis	To assess positive and negative economic and physical effects independently or support simulation and optimization models for systems analysis	Well-defined cost–benefit models may translate environmental aspects into economic terms. However, the intergeneration externalities are very difficult to address.
Optimization model	To reach the best solution among numerous alternatives, considering one or several objectives.	Models have solved the following issues: <ul style="list-style-type: none"> • single network planning (Anderson and Nigam, 1967; Anderson, 1968; Fuertes et al., 1974; Helms and Clark, 1974; Kuhner and Harrington, 1975; Jenkins, 1979; Clayton, 1976; Rao, 1975) • dynamic, multi-period investment (Marks et al., 1970; Marks and Liebman, 1971) • size and site facilities (Chapman and Yakowitz, 1984; Li and Huang, 2006a,b, 2009a,b; Nie et al., 2007; Li et al., 2007, 2006, 2008a,b; Huang et al., 2001, 2002; Xu et al., 2009) • manage infrastructures like landfill (Davila et al., 2005)
Simulation model	To trace the lengthy chains of continuous or discrete events based on cause-and-effect relations describing the operations in complex systems and helping investigate the dynamic behavior of the system (Wang et al., 1996).	Models developed: WRAP (USEPA, 1977) Models developed for SWM systems: SWIM (Wang et al., 1996), GIGO (Lawver et al., 1990; Anex et al., 1996), AWAST (Villeneuve et al., 2009), EcoSolver IP-SSK (Krivtsov et al., 2004), TASAR (Tanskanen and Melanen, 1999)
Forecasting model	To characterize waste streams quantitatively and qualitatively and construct a management information system to accumulate information over time. To predict waste generation, time-series regression analysis (Katsamaki et al., 1998; Navarro-Ésbri et al., 2002), system dynamics models (Dyson and Chang, 2005), and other regression models have been applied (Grossman et al., 1974).	Models have related variables like population (Grossman et al., 1974), income level (Grossman et al., 1974; Beigl et al., 2005), dwelling unit size (Grossman et al., 1974), total consumer expenditure and gross domestic product (Daskalopoulos et al., 1998), production measures, household size, age structure, health indicators (Beigl et al., 2005), per capita retail and tipping fees for waste disposal (Hockett et al., 1995) to waste generation, total income per service centre, people per household, historical amount generated, income per house and population (Dyson and Chang, 2005).
Integrated modeling systems	To improve synergistic connections among different models, concatenating their total functionalities.	IMS have provided: <ul style="list-style-type: none"> • dynamic information of waste generation and waste shipping (Chang et al., 1993) • optimal capacity expansion patterns for waste-to-energy and landfill facilities over time (Baetz, 1990) • Models developed: ORWARE (Dalemo et al., 1997; Björklund et al., 1999)

4. Systems analysis used for solid waste management in European countries

4.1. Methodology

Several types of SWM systems in European countries can be identified and classified. The characterization of these systems in the EU and its member states was mainly performed by the authors in 2008 and 2009 based on the databases developed by European Topic Centre on Sustainable Consumption and Production. In the case of Belgium and Spain, for example, the inquiry was made through the regional entities of Flanders and Catalonia, respectively. To understand which are the research needs and future directions in regard to the systems analysis techniques for SWM in European countries, a comparative analysis was also conducted in this paper for the distinction of relevant applications.

4.2. Waste management systems in European countries

From a life-cycle point of view, an all-inclusive MSW management system includes all essential operational units from collection, to shipping, to treatment, to recycling, and to disposal. Yet the current European regulations promoting the hierarchy of waste management inevitably involve a wealth of waste management

practices tied to policies, institutional settings, financial mechanisms, technology selection, and stakeholder participation. For instance, the landfill directive promoted biodegradable municipal waste (BMW) management systems, which focus on building separate collection systems provided by local authorities through specific bins leading to separate, mandatory BMW treatment systems (Austria, Netherlands). Some of the EU member states applied economic instruments including Pay-As-You-Throw (PAYT) and an organic waste tax to create economic incentives for residents to divert BMW from regular waste streams normally being collected by municipalities to specific collection avenues. They recognized that the diversion costs that each waste disposal authority would face would differ according to the particular circumstances (EIONET, 2007a). For example, both BMW system and Landfill Allowance Trading System (LATS) in the United Kingdom (UK) were launched to provide local authorities with the flexibility to manage waste streams more effectively. The LATS system revolves around transferable allowances which enable the greatest amount of waste diversion to occur in areas where it is cheapest, and most practicable to do so.

The Packaging waste directive has also promoted similar incentives by using the “EPR system” and the “deposit-refund system” to ensure the maximum reuse and recycling. The most well known EPR system is the packaging waste Duales System

Table 2

The contribution of systems assessment tools to SWM systems (Chang et al., in press).

Systems assessment tools	Description	Contribution to SWM systems
Management information system, decision support system and expert systems	Consists of different methods applied to exchange and manage information; used to help in decision making	MIS/DSS/ES have been applied: <ul style="list-style-type: none"> to provide information storage and transmission through countries (EIONET, 2009) to yield specific decision support (Chang and Wang, 1996; Barlishen and Baetz, 1996; Hastrup et al., 1998; Bhargava and Tettelbach, 1997; AEA Technology, 1998) to relate waste stream characterization with implications on shipping, processing and disposal of waste streams (MacDonald, 1996)
Scenario development	To create hypothetical sequences of events constructed for the purpose of focusing attention on causal processes and decision points (Kahn and Wiener, 1967)	Has the ability to explore events (events in this case are policies and decisions taken) that might occur associated with SWM on a temporal scale. Such events can be inside or outside the SWM system. Fell and Fletcher (2007) have contributed with scenario developments for future lifestyle trends and forecasting based on lifestyle scenarios for waste composition
Material flow analysis	Consists of a systematic assessment of the flows and stocks of materials within a system defined in space and time (Brunner and Rechberger, 2003)	Software developed in MFA: SFINX (van der Voet, 1995a,b), FLUX (Huijbregts, 2000), STAN (TU Vienna, 2009), DYNFLOW (Elshkaki, 2000), GaBi (PE International, 2006) and Umberto (IFU, 2006)
Life cycle assessment	Consists of a process to evaluate environmental burdens associated with a product, process or activity by identifying and quantifying energy and materials used, wastes and emissions released to the environment, to assess impact of those energy and material uses and releases and to identify and evaluate opportunities that lead to environmental improvements (EEA, 2003)	Models developed for SWM systems: IWM (White et al., 1995; McDougall et al., 2001), WASTED (Diaz and Warith, 2006), WISARD/WRATE (Ecobilan, 2004; Buttol et al., 2007), EASEWASTE (Christensen et al., 2007)
Risk assessment	To relate environmental and human health risk to accidents quantitatively, through a statistical evaluation	Help in the evaluation of transversal SWM systems
Environmental impact assessment	A procedure that aims to ensure that the decision-making process concerning activities that may have a significant influence on the environment takes into account the environmental aspects related to the decision (Tukker, 2000)	EIA associated to a specific project attempts to solve controversial issues from the target project such as siting issues originated from the NIMBY effect, technical issues to justifying the choice of technology for emission reduction, and even the rejection of the project (Chang et al., 2009). In Europe, EIA is mandatory for landfills and incineration plants with regard to capacity limits through EU Directive 85/337/EEC (EU, 1985), as amended by EU Directive 97/11/EC (EU, 1997). A good example can be found in Barker and Wood (1999). Its applicability is emphasized by EU Directive 2001/42/EC (EU, 2001), to which it is obligated for the promotion and elaboration of an SEA for SWM plans. More details can be found out in Dutch Ten Year Program on Waste management 1992 and 2002 (Verheem, 1999)
Strategic environmental assessment	Consists of the environmental assessment of a strategic action as a policy, a plan or a program (Thérivel and Partidário, 1999)	Has allowed the inclusion of user-charges, landfill disposal fees, recycling credits, product charges, deposit-refund schemes, and producer-responsibility schemes into the decision making in SWM systems, promoting a more sustainable management of waste. For such purposes, several methodologies have been applied: CBA-based LP (Chang et al., 1997, 1996), CBA-based MIP (Chang et al., 2005), CBA-based fuzzy goal programming (Chang and Wang, 1997), fuzzy contingent valuation (Chang et al., 2009), minimax regret optimization (Chang and Davila, 2007), GIP-based game theory (Davila et al., 2005), CBA-based MCDM (Karagiannidis and Moussiopoulos, 1997; Rousis et al., 2008), optimal control of landfill space (Chang and Schuler, 1991) inexact fuzzy-stochastic constraint (Li et al., 2009), IOA (Brahms and Schwitters, 1985; Franklin Associates, 1999; Gay et al., 1993; Hekkert et al., 2000; Joosten et al., 2000; Patel et al., 1998; Nakamura, 1999; Pimenteira et al., 2005)
Socioeconomic assessment	Consists of computer-based practices that apply integrated market-based and/or policy/regulation requirements for SWM	SWM systems assessed to reach sustainable management, focusing on different aspects. Models developed: LCA-IWM (den Boer et al., 2007) and MSW-DST (Thorneloe et al., 2007; Weitz et al., 1999). Several methods have been combined to reach sustainability: Cherubini et al. (2008) have combined LCA with MFA and energy analysis methods, Nakamura and Kondo (2002) used IOA and LCA to construct a waste input–output model, Huppel et al. (2006) and Tukker et al. (2009) have combined both methods to obtain IOA with environmental extensions for different sections (including waste management sectors). A Geographical Information System (GIS) combined with LCI, EIA and optimization model has been promoted by Chang et al. (2008, 2009) for landfill siting
Sustainable assessment	Refers to the integration of different methodologies in such a way that obtaining an analysis, an evaluation or a planning that approaches several management aspects in which sustainability implications may be emphasized and illuminated	

Deutschland (DSD) (or Green Dot system) that was firstly applied in Germany in the 1990s and later on all over Europe (Buclet, 2002). The basic idea of the DSD is to establish a privately organized channel assuring that all primary packaging can be collected from the consumers will then undergo a material-specific recycling process through the consumers and service providers. This is done through the so called “Green Dot” which is a label on packaging material used to identify the product belonging to the dual system during the consumption phase (Klepper and Michaelis, 1994). As for these deposit-refund systems, Dansk Retursystem is one of the oldest ones, and has been in use since 1984; it is applicable for refillable, non-refillable, reusable and disposal, and ready-to-drink beverages and mineral water bottles (Pro-Europe, 2009). However, in Denmark there is no producer-responsibility scheme, namely no separate management system for packaging waste (Danish EPA, 1999; Pro-Europe, 2009), making the costs for handling packaging waste uncertain, and resulting in a higher budget for waste management in local authorities (EEA, 2005).

The remaining part of waste streams usually called residual household waste or mixed municipal waste still need to be cleaned up by municipalities and local authorities. The Danish Waste Model is a representative residual/mixed waste system as it is based on a joint venture to form a coherent whole (Danish EPA, 2001). According to the Danish EPA (2001), the structure of the SWM systems is characterized by the following principles: 1) the system includes all types of waste (e.g. household, industrial and hazardous waste); 2) the responsibility for the SWM lies solely with the local authorities (council), which are responsible for establishing capacity for waste management and for providing information on how to dispose of the waste produced within the local council, irrespective of whether this waste originates at households or trade and industry (EIONET, 2007b); 3) the duty to assign waste treatment and disposal facilities lies with the local authorities, and waste generators are bound to those who use them; 4) financing of the system rests on the polluter-pays principle (PPP); and 5) waste collection and waste treatment rest on the principle of source separation (EIONET, 2007b). With these principles, the systems boundaries are quite well defined, such as packaging waste collection created by local authorities was managed by an external system. Considering the main waste streams in MSW – residual waste, BMW and packaging waste – a review on SWM systems in European countries may be summarized in Table 3.

In addition to existing SWM systems handling related material and cash flows at scales, there is also a need for building up proper information exchange platforms capable of offering decision makers the basis for the assessment of relevant projects, programs, and plans. Member states must provide information to supply European waste information systems like Eurostat, EIONET, and ReportNet. Eurostat's main role is to process and publish comparable statistical information at European level (Eurostat, 2009). The Eurostat data centre on waste is responsible for providing robust data, indicators and other relevant information for assessing the effectiveness of the community waste policy (Eurostat, 2009). The functions of the Eurostat information system are divided into four sectors which correspond to the various stages in the processing of data from their collection to their dissemination including production (collection, validation and storage of the data and meta-data), storage of the reference data (acceptance of the information), use of the reference data (visibility/security and find/deliver), and dissemination of information (Dubois, 1997).

EIONET is a collaborative network of the European Environment Agency (EEA) and its member states, connecting National Focal Points in EU and accession countries, European Topic Centres, National Reference Centres, and Main Component Elements. EIONET provides a mechanism whereby National Focal Points in

European countries can make documents available to the EEA and also retrieve documents of interest from the EEA. The integrated electronic workplace environment allows online collaboration between environmental personnel across Europe. EIONET supports the collaborative process and reduces the reporting burden for environmental protection agencies across Europe (EEA, 2002).

ReportNet aims to develop common tools and a shared information infrastructure as the European Environmental Information System; and it is based on a set of inter-related tools and processes which all build on the active use of the World Wide Web (EIONET, 2009). ReportNet is EIONET's infrastructure for supporting and improving data and information flows (EIONET, 2009). With this platform, ReportNet also aims at providing an effective network infrastructure whereby collaboration can be achieved so that the environmental reporting burden of the EU member states can also be reduced (EEA, 2002).

4.3. Comparative analysis

To understand how to deal with multiple alternatives and a plethora of outcomes regarding systems analysis for SWM in European countries, a comparative analysis was conducted based on 218 applications in this paper. Such a comparative analysis is presented in Fig. 3. Sometimes the assessments conducted for waste management overlap with several SWM systems simultaneously; such cases happened when a specific decision maker responsible for waste management at a geographical area of interest is the same as the decision maker inside such boundaries handling several sub-systems.

Fig. 3 clearly indicates that 1) more cases were associated with MSW, 2) the studies on residual/mixed waste and packaging waste streams were received equal emphasis, and 3) BMW received the least attention. Such phenomena can be explained by the fact that the BMW system is less established in European countries, as opposed to other parts of the world. Comparing the relative distribution between groups of models and tools for systems analysis, the most common practices for waste management in European countries are those using various systems assessment tools rather than system engineering models. Table 4 further confirms the same observations after the realization of application metrics of systems assessment tools.

The comparison across the boundaries implies that systems assessment tools have been applied to evaluate and help in decision making based on environmental issues have great potential to integrate other aspects, like economics or social impacts. Given that EU regulations have given emphasis to EIA in SWM, the inclusion of EIA has become favored by decision makers at national, regional, and local levels. Thus, the primary stage of decision analysis normally leads to assess a suite of management options, evaluate managerial and strategic plans, and collect and share information. Likewise, climate change and resources depletion are emerging issues of most concern, which are even more influential when managing SWM, and decisions and policies were oftentimes made with the aid of LCA or LCI in public institutions.

With fewer applications, systems engineering models were capable of studying waste production processes and assessing the interactions in numerous types of SWM systems addressing impacts from technical to social, and to economic perspectives. However, such applications are not easy to implement since the necessary assumptions being made may or may not be realistic. As a consequence, systems engineering models have not been applied to the same extent as systems assessment tools in EU member states. Oftentimes, these models are not geared toward helping decision makers' needs. Their contribution is often limited to use a mathematical functional form structured to derive strategic

Table 3
Waste management systems in European countries.

Country	Residual/mixed waste system	BMW system	Packaging system
Austria	Yes, with ban of landfill regulation for BMW	Yes, being mandatory with penalties	<ul style="list-style-type: none"> • Alstoff Recycling Austria system (Green Dot Dystem) – EPR • Bonus Holsystem (commercial packaging waste) – EPR • Öko-Box for beverage carton containers • Pet2Pet for polyethylene terephthalate (PET) bottle • Deposit-refund system for beverage containers but only mandatory for refillable plastic beverage containers • ARO system for wastepaper • AGR system for glass • ArgeV for lightweight fraction • Fost Plus (Green Dot System) – EPR
Belgium	Yes, with cash tax, residual waste and/or environmental tax and organizations like Fost Plus	Yes, with PAYT	
Denmark	Yes, with collection fee based on polluter-pays principle	Yes, but is voluntary and only for garden waste	<ul style="list-style-type: none"> • Dansk Retursystem – deposit-refund system • Packaging glass system: recycling schemes • Wastepaper and waste cardboard system: recycling schemes are mandatory • Packaging tax
Finland	Yes, with waste charge: sorted materials pays less	Yes, promoted by information instruments	<ul style="list-style-type: none"> • Suomen Palautuspakkaus Oy (Palpa) – deposit-refund system for packaging waste, including beverages • Newspaper, copy paper and other paper products and packaging waste – EPR • Beverage containers – packaging tax (exemption or lower tax rates only if package is part of a returnable deposit scheme)
France	Yes, with fees	Exists but not consolidated	<ul style="list-style-type: none"> • Eco-Emballages (Green Dot System) – EPR
Germany	Yes	Bio-Bin system and other mandatory systems	<ul style="list-style-type: none"> • Duales System Deutschland – EPR • Deposit-refund system
Greece	Yes, with fees to cover the service	No	<ul style="list-style-type: none"> • Green Dot System called HERRCo – EPR • KEPED – packaging waste system for waste oils • Supermarkets as individual systems
Ireland	Yes, with PAYT	Exists but not consolidated	<ul style="list-style-type: none"> • Repak – Green Dot system – EPR
Italy	Yes, with municipal waste tariff	Yes	<ul style="list-style-type: none"> • Consorzio Nazionale Imballaggi (CONAI) (Green Dot System) – EPR
Luxembourg	Yes. With PAYT system	Yes (green bins)	<ul style="list-style-type: none"> • Valorlux (Green Dot System) – EPR
Netherlands	Yes, with levy	Yes, mandatory	<ul style="list-style-type: none"> • Compulsory deposit return systems for refillable glass bottles and one-way packaging • Nedvang – EPR • Stichting Retourverpakking Nederland – one-way deposit-bearing PET containers for soft drinks and water larger than 0.5 L • Wastepaper and waste cardboard – EPR • Norsk Resy AS – packaging waste and corrugated and solid board packaging – EPR • Norsk GlassGjenvinning AS for glass • Norsk MetallGjenvinning AS for metal • Gront Punkt Norge for plastic packaging, beverage cartons and carton packaging • Norsk Resirk AS – deposit-refund system for beverage packaging, steel and aluminium cans, plastic bottles non-refillable • Sociedade Ponto Verde (Green Dot System) – EPR • Valormed (Medicine packaging waste) – EPR • Marão mineral water system (private) – deposit-refund system for one-way PET bottles of Marão trend mark • Ecoembes SL (Green Dot System) – EPR • Ecovidrio, for glass packaging – EPR
Norway	Yes, with waste tariffs	Yes, with organic waste tax	
Portugal	Yes, by water consumption fee	No	<ul style="list-style-type: none"> • Sociedade Ponto Verde (Green Dot System) – EPR • Valormed (Medicine packaging waste) – EPR • Marão mineral water system (private) – deposit-refund system for one-way PET bottles of Marão trend mark • Ecoembes SL (Green Dot System) – EPR • Ecovidrio, for glass packaging – EPR
Spain	Yes for Catalonia, with a landfill tax, incineration tax	Only in Catalonia for municipalities with >5000 inhabitants	
Sweden	Yes	Yes	<ul style="list-style-type: none"> • EPR for several waste streams like packaging and wastepaper • Deposit-refund systems for cans, plastics and glass bottles • Returpack – deposit-refund system for all plastic and metal beverage containers for ready-to-drink beverages, including refillable glass bottles • Beverage bottles – EPR • Reusable packaging – deposit-refund system • PET-Recycling Schweiz – packaging PET one-way – EPR • IGORA for aluminium cans – EPR • Ferro-Recycling for tinfoil – EPR • VetroSwiss for glass – EPR, mandatory system • Disposable packaging in PVC – obligatory deposit • Wastepaper and cardboard system related to municipalities/local authorities • PRN system – EPR • PERN system – EPR
Switzerland	Yes, with Canton tax	Yes, whenever possible	
United Kingdom	Yes, landfill tax only for companies, local authorities or other organization	Yes, LATS and for garden waste (from civic amenity)	

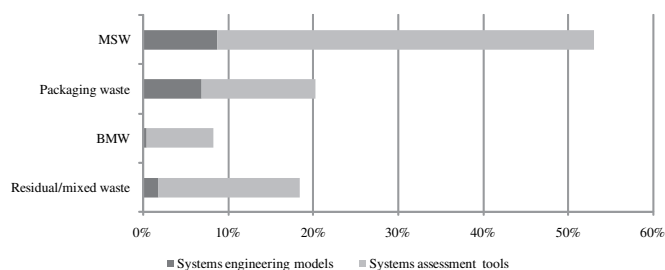


Fig. 3. Systems analysis applied in solid waste management systems in Europe.

guidelines and/or orientations in an SWM system. Sometimes, the mathematical outputs are contradictory with existing ideas that have already embedded in decision makers' minds. In such cases, CBA may be defined or refined well to fit in the LCA framework in the decision making arena, the application potential may be improved. Table 5 further confirms the observations across MSW, residual/mixed waste, packaging waste, and BMW after the realization of application metrics of systems assessment tools.

Table 6 presents individual efforts in each country level; apparently all European countries involved can be classified into three groups according to their application of systems analysis methodology. Countries that have mostly applied systems analysis techniques include Italy, Sweden, the United Kingdom, and Denmark; countries with a moderate number of applications include France, Germany, Austria and Finland; and countries with low interest in such applications include Spain, Greece, Ireland, Luxembourg, Portugal, Belgium, the Netherlands, Norway and Switzerland. Why do such discrepancies occur, given that all EU member states are under the same European guidelines? It is mainly due to national differences in waste management policies within each country. In countries where more systems analysis practices have been applied, the driving forces for the applications are related to the need for solving multi-faceted environmental issues linked with various facilities while complying with international, EU, regional, and local regulations.

Sweden, a country where sustainability principles were heavily promoted by national/international entities and its waste management policies were born based on EPR, can be selected as a particular case. Such sustainability principles have infiltrated into the municipality level, such as Stockholm, which is the city where the development of models to aid in decision making of SWM was favored. In Denmark, the packaging refundable system was assessed due to the presence of the Packaging Waste Directive. In Italy, due to a legal action against it within EC in a waste crisis that had plagued Naples and the Campania region for more than ten years, Mastellone et al. (2009) conducted an MFA to provide scientific support for decision makers who were managing the waste crisis. These three cases, including Sweden, Italy, and Denmark, concur with our observations.

Concerning countries which have applied systems analysis techniques only moderately, motivations for doing so are based on the political concern with regard to environmental impacts due to SWM. For example, in France, Ecobilan-Pricewaterhouse Coopers developed several assessments for Agence de l'Environnement et de la Maîtrise de l'Energie (ADEME) concerning SWM. Additional motivations were tied to the need to comply with European regulations and to manage the waste streams in a sustainable way. As for those European countries with no prevalent applications of systems analysis techniques, the true reason is that they were lacking sustainable development concepts and communication channels among stakeholders in the waste management regime.

Even though some European countries did not promote systems analysis practices for SWM solely, it did not imply that the issues of SWM were ignored. LCI/LCA was deemed the most popular system assessment tool in the EU so far. In the nexus of environmental management and industrial ecology, LCA is actually a normalized method in connection with the norm family ISO 14040 (2006). In reality, LCA may be integrated with other system assessment tools while standing up to close scrutiny and achieving a higher level of evaluation for SWM system wide. In addition, further advantages via applying LCA/LCI can be assured by the elaboration of environmental indicators in the scope of ISO 14025 (2006) (Gallo et al., 2009). Both ISO 14040 and ISO 14025 allow the gap to be bridged between waste management and industrial ecology. For example, at this juncture, the Netherlands applied LCA, MFA, CBA and EIA collectively to materials management, which includes product use phase and waste production phase.

SA models, mainly developed in Sweden, were the second mostly applied system assessment tool in the EU for SWM because of the possible linkage between SWM and energy recovery (e.g., waste-to-energy) given the presence of a large number of incineration plants. Besides, the ORganic WASTE REsearch (ORWARE) model, mainly developed in Sweden, combines the concept of LCA and MFA to simulate and assess MSW and BMW systems. This type of system analysis that is considered highly novel in Sweden has been widely applied. On the other hand, the application of MIS/DSS/ES, the third mostly applied systems assessment tool in the EU, is related to the need to provide information flows among the EU member states and to evaluate how member states carry out legislative measures. Such a mandatory process was defined simultaneously by several European Directives and Regulations, like EU 91/692/EEC (EU, 1991), EU 2003/35/EC (EU, 2003), and Waste Statistics Regulation no 2150/2002 (EU, 2002). Almost all European countries had already developed MISs at regional and national levels, like in the North Rhine-Westphalia region in Germany, Italy, Denmark (ISAG information system), United Kingdom, and Austria. Such MISs may also improve connections of waste producers and consumers, which have been channeled for possible waste exchange activities almost all over Europe. Many European projects in this direction have been developed. A salient case co-developed by Austria, Belgium, Germany, and the Netherlands is the EUDIN (European Data Interchange of Waste Notification System) which is an electronic data interchange platform for waste transportation control within, into and out of Europe boundaries. It consists of an electronic data base that enables an electronic exchange of the data of the notification form and the movement/tracking form (EUDIN, 2002). Such an information sharing platform can be further integrated and improved to develop more powerful ESs and DSSs for waste management. Further justification for the use of MIS/DSS/ES has been for the siting of infrastructures like landfills and incineration plants since some systems engineering models for siting infrastructures can be handled by GIS to make spatial interactions comprehensive and understandable for decision makers instead of domain experts in SWM systems (Chang and Wang, 1996; Barlishen and Baetz, 1996; MacDonald, 1996; Haastrup et al., 1998).

5. Future perspectives of systems analysis for solid waste management in Europe

5.1. Current status and limitations

The assessment of SWM by using systems analysis techniques allows decision makers to learn about total system complexity. A quantifying complexity factor requires evaluating interfaces. Whereas system assessment tools provide a wealth of composite

Table 4
The purposes of systems assessment tools.

SWM systems	Systems assessment tools	References
MSW	To collect and share information flows in SWM	Nationale Reststoffenbeurs, 1986; Waste Exchange UK, 2000; CIWM, 2003; Dall et al., 2003; Becker et al., 2007; Denmark Waste Exchange, 2008; LUA NRW, 2006; International Synergies Limited, 2007; Fahy, 2007; Economie, 2008; Jean-Gerard, 2008; APA, 2008; IHK Recyclingbörse, 2008; IWEN, 2008; EIONET, 2009; Mochty, 2009
	To understand environmental impacts in SWM system with respect to pollutant fate and transport, or the waste itself	Dahlbo and Assmuth, 1997; Obernosterer and Brunner, 1997; Powell et al., 1996, 1999; Döberl et al., 2002; Melloni et al., 2003; Bolze, 2004; Beigl and Salhofer, 2004; Sokka et al., 2004; Ecobilan, 2004; Xará et al., 2005; Kirkeby et al., 2005; Badino et al., 2007; Mastellone et al., 2009; Rigamonti et al., 2009a,b; Frakgou et al., 2009
	To assess SWM plans, regulations, policies, and strategies	EU, 1997; Björklund et al., 1999; Saarikoski, 2000; Arbter, 2001; Moberg et al., 2002; Aumônier, 2002; Ministry of the Environment Government of Japan, 2003; Salhofer et al., 2005, 2007; Pladerer et al., 2007; SEA Wiki, 2007; Escalante et al., 2007; Buttol et al., 2007; Cheshire County Council, 2007; SEPA, 2007b; Pisoni et al., 2009; NLWA, 2009; Desmond, 2009
	To assess options for decision making in SWM systems	Sundberg, 1993; Karagiannidis and Moussiopoulos, 1997; Sivertun and Le Duc, 1998; Ljunggren, 1998, 2000; Wilson, 2002; Reich, 2002; Fiorucci et al., 2003; Karagiannidis et al., 2003; Skordilis, 2004; Muñoz et al., 2004; Costi et al., 2004; Viotti et al., 2005; Eriksson et al., 2005; Reich, 2005; Gentil et al., 2005; Dornburg and Faaij, 2006; Jansen and Gerlo, 2006; Minciardi et al., 2007; SEPA, 2007a; Ulli-Beer et al., 2007; Bovea and Powell, 2006; Rodríguez-Iglesias et al., 2007; Cherubini et al., 2008; Gallo et al., 2009; de Feo and Malvano, 2009; Federico et al., 2009; Ekvall et al., 2009; Tunesi and Rydin, 2009; Abeliotis et al., 2009
	To site infrastructures To assess part of the system including waste production steps with respect to perspectives	Lahdelma et al., 2002; EEA, 2003 Finnveden et al., 2002; Fell and Fletcher, 2007; Bovea et al., 2007; Grosso et al., 2008; Karadimas and Loumos, 2008
Residual/ mixed waste	To collect and share information flows in SWM	LUA NRW, 2006
	To assess environmental impacts related to SWM infrastructures	Harrop and Pollard, 1998; Coutinho et al., 1998; Snary, 2002; Allgaier and Stegmann, 2003; Verro et al., 2003; Capuzzo and Farina, 2003; Cossu et al., 2003; Boerboom et al., 2003; Marques and Hogland, 2003; Belgioirio et al., 2003; Belfiore et al., 2005; Zorzi et al., 2005; Morra et al., 2005, 2006; Masi et al., 2007; Bour and Zdanevitch, 2007; Cangialosi et al., 2008; Moutavtchi et al., 2008; Perkoulidis et al., 2010
	To assess SWM options and the system itself	Loeschau and Rotter, 2005; van der Linden and Torfs, 2005; Chanchampee and Rotter, 2007
	To evaluate operations occurring in the SWM system (collection, treatment, and disposal)	Ecobilan, 2004; Bergsdal et al., 2005; Emery et al., 2007; Wittmaier et al., 2009
	To site infrastructures To assess policies and economic instruments	Vaccari et al., 2005 Nilsson et al., 2005; Björklund and Finnveden, 2007
Packaging waste	To collect and share information	GS1, 2008
	To assess management options for a specific packaging material, considering environmental perspectives	Finnveden et al., 1994; Kaila, 1998; Ryberg et al., 1998; Person et al., 1998a,b; Widheden et al., 1998a,b; Frees et al., 1998, 2004; Detzel et al., 2003; Pancaldi et al., 2005; Schmidt et al., 2007; Dahlbo et al., 2007
	To assess management options for a specific packaging material considering targets to be established	Dalager et al., 1995; Fehrer and Brunner, 1997
	To analyze the specific parts of the system, including collection, treatment, and disposal	Baumann et al., 1993; Finnveden and Ekvall, 1998; Holmquist, 1999; Rutegård, 1999; Ibenholt and Lindhjem, 2003; Ecobilan, 2004
	To assess and compare different SWM systems applied to a specific packaging waste	Frees and Weidema, 1998; Ekvall et al., 1998; Jahre, 1998; Ekvall and Bäckman, 2002; Hirschier et al., 2005; Heilmann and Winkler, 2005; Dahlbo et al., 2005; Vercauteren et al., 2007
BMW	To assess policies	Bruvoll, 1998; Wäger et al., 2001
	To assess and improve the system, including environmental perspectives	Björklund et al., 2000; Wassermann et al., 2003; Shmelev and Powell, 2006; Güereca et al., 2006; Schmidt and Pahl-Wostl, 2007; EUNOMIA, 2007
	To understand environmental impacts in SWM system with respect to pollutant fate and transport, or the waste itself	Boldrin and Christensen, 2007
	To understand the source of the waste streams	Purcell and Magette, 2007, 2009
	To compare system outputs with substitute products	Eriksson et al., 2002
	To compare technologies applied to the collection, treatment and disposal in SWM systems	Edelmann and Schleiss, 1999; Danish EPA, 2003; Lang et al., 2006a,b

measures of complexity inside procedures/components and between them, joint formulation of human factors and physical/biochemical features in system engineering models in concert with those well-defined procedures/components brings about a considerable contribution to the improvement of SWM. Such a successful joint endeavor across the boundaries between models and tools can

be evidenced by the development of EUDIN, LATS, Green Dot system, and ORWARE.

Without good practices or guidelines, however, such joint endeavors inevitably tend toward the cost-effectiveness principle. For example, a potential social factor that can drive the implementation of systems analysis toward a cost-ineffective condition is

Table 5

The purposes of systems engineering models.

SWM systems	Systems engineering models	References
MSW	To predict solid waste production	Brahms and Schwitters, 1985; Dennison et al., 1996a,b; Andersen et al., 1998, Patel et al., 1998; EEA, 1999; Navarro-Esbrí et al., 2002; Lebersorger et al., 2003; Beigl and Lebersorger, 2009
	To optimize the system for choosing the best option	Kaila, 1987; Hokkanen and Salminen, 1997; Gottinger, 1988; Cosmi et al., 1998; Komilis, 2007
	To assess recycling rate	Huhtala, 1997
	To site infrastructures	Mitropoulos et al., 2009
	To analyze specific parts of the system	Tanskanen and Melanen 1999; Villeneuve et al., 2005, 2009
	To assess the system	MCCK and Consultancy, 1998
Residual/mixed waste	To site infrastructures	Arnold and Terra, 2006
Packaging waste	To analyze how to reach recycling targets	Radetzki, 1999; Angst et al., 2001
	To study/predict waste production	Bach et al., 2004; Maunder et al., 2006
	To analyze the specific parts of the system, like collection, treatment disposal	Hanley and Slark, 1994; Powell et al., 1995; Tucker et al., 1998; Wäger et al., 1998; Ekvall and Bäckman, 2001; Petersen and Andersen, 2002
	To understand and know social cost and benefits of different packaging waste systems	Vigso, 2004
	To assess policies	McHenry et al., 2003
BMW	To assess the system	Le Bozec et al., 2009

the NIMBY (Not in my backyard) syndrome. More human factors that may tilt the balance/scales of the SWM system should be certainly included to address the impacts from socioeconomic conditions, policy instruments, and regulatory requirements such as EU Directives, national regulations, and regional or local plans and strategies. With the aid of the technology hub proposed in Fig. 1, which pinpoints the synergistic effect between system engineering models and system assessment tools, the SWM communities become able to get over the hurdle of system complexity to some extent. A salient case of regulatory requirement is the mandatory EIA and SEA for some specific cases due to the emergency of European Directives 85/337/EEC (EU, 1985) and 2001/42/EC (EU, 2001), respectively. While European Directives with mandatory targets and features are significant, national policy instruments may drive more incentives to achieve the prescribed goals by a more cost-effective, efficient, and forward-looking way. In this pathway, CBA, LCA, MFA, and others may be glued together to support high-end analysis.

5.2. Gaps on knowledge of waste management

All of these system complexities may encourage the creation of a system of systems (SoSs), which may include large-scale concurrent and distributed sub-systems in relation to the WHP. In

other words, each SoS might be a collection of task-oriented or dedicated sub-systems that pool their resources and capabilities together to obtain a more specific goal from an integrated solid waste management perspective. The selected system assessment tools in support of developed scenarios are the workhorses that may enrich the SoS and empower the systems analysis when dealing with contemporary, emerging challenges. These challenges include but are not limited to climate change, resource depletion, and energy crisis as they are the long-term challenges facing SWM communities. Systems engineering models in concert with system assessment tools may be capable of contributing to a fundamental understanding of environmental, technical, economic and social aspects of SWM systems in response to these challenges. To quantify the pros and cons of each alternative at the EU or national level, however, green accounting might be an additional tool for tackling these challenges at different scales.

With the increase of stakeholders' involvement, information flows in and out of the SWM systems increase the complexity. Additionally missing data and information in terms of both quantity and quality can make such high-end systems analysis difficult to advance. Thus, the need to effectively and efficiently collect data through identifiable data sources, recognized pathways, and involved agencies, may be justified with respect to the related complexity via the development of MISs, ESs and DSSs. The need for

Table 6

Number of published articles studying SWM systems in European countries.

Countries		AT	BE	DE	DK	ES	FI	FR	GR	IE	IT	LU	NL	PT	SE	UK	NO	CH	Total
Systems engineering models	CBA	1	0	0	2	0	0	1	0	1	0	0	0	0	2	3	2	0	12
	FM	4	0	0	1	1	0	0	1	2	0	0	0	0	0	1	0	0	10
	SM	0	0	0	0	0	1	3	0	0	0	0	0	0	0	1	1	1	7
	OM	0	0	1	0	0	2	0	3	0	1	0	0	0	0	0	0	0	7
	IMS	0	0	0	0	0	1	0	0	0	0	0	0	0	0	1	0	0	2
Systems assessment tools	MIS/DSS/ES	2	1	7	2	0	0	1	1	1	4	1	1	1	1	3	0	0	26
	SD	0	0	0	0	0	0	0	0	0	0	0	1	0	0	1	0	0	2
	MFA	2	1	2	0	1	2	0	0	0	1	0	0	0	0	0	0	1	10
	LCA/LCI	1	2	7	13	6	2	13	0	0	10	0	2	1	7	5	1	2	72
	RA	1	0	0	0	0	0	1	0	0	7	0	1	0	1	2	0	0	13
	EIA	0	0	0	0	0	1	0	0	0	0	0	0	1	0	0	0	0	2
	SEA	4	0	0	0	0	1	0	0	1	1	0	2	0	0	3	0	0	12
	SoEA	0	0	2	0	0	0	0	0	0	3	0	0	0	4	1	0	2	12
	SA	1	0	0	1	0	5	0	3	2	5	0	1	0	10	2	0	1	31
	Total	16	4	19	19	8	15	19	8	7	32	1	8	3	25	23	4	7	218

Note: Country abbreviations – AT (Austria), BE (Belgium), DE (Germany), DK (Denmark), FI (Finland), FR (France), GR (Greece), IE (Ireland), IT (Italy), LU (Luxembourg), NL (Netherlands), NO (Norway), PT (Portugal), CH (Switzerland), SE (Sweden), ES (Spain), UK (United Kingdom). Bold numbers represent meaningful countries concerning the number of published articles in systems analysis (more significant in systems assessment tools).

a Quality Assurance System (QAS) to enhance confidence in data and information products can be assured. Such electronic platforms (e.g., MIS, ES, and DSS) would be helpful to support decisions and facilitate research to gain deeper knowledge in SWM. Only this way will it become possible to reach the end-of-waste criteria, supporting waste recycling, treatment and disposal and create a more sustainable management.

5.3. Research needs for the future

SDS is a precautionary principle and European directives have reflected it to some extent. However, waste prevention programs at the EU level should have an important role in system analysis because failures can occur in association with environmental, economic and social aspects. These waste prevention plans in SWM systems generally have multi-objective, interactive, dynamic, and uncertain features that complicate the applications of modeling and assessment techniques. The application of EPR, such as the Green Dot system, should be expanded as an integral part of modern SWM systems, since it may provide a possible route to maximize resources utilization and confirm the sustainability. To achieve this goal, carrying out site-specific and process specific CBA, MFA, LCA, EIA, etc would be required. With these site-specific and process specific inputs, next-generation systems engineering models would be able to reflect environmental impacts through an integrated approach. It should lead to consider more options across waste treatment technologies at all planning, construction, and operational stages, and evaluate more policy instruments to promote waste prevention, reuse and recycling.

To achieve these high-end decision analyses, systems engineering models may be simultaneously and flexibly integrated with system assessment tools in the context of IMS or may be sequentially applied in multiple stages so that the results from one model or tool are the inputs needed for the next one. Given that the New Directive of Waste defines public participation in the assessment of waste management plans and waste prevention programs through SEA, conducting such quantitative decision analyses should include more stakeholders in the decision making process. With this trend, future CBA, LCA, MFA, EIA and SEA might become a multitude of essential models and/or tools that may be mandatory in specific situations. On many occasions, we envision that LCA should be designed based on the framework of MFA since the object to be assessed is a process, not a product. Besides, an MIS would be essential to manage information flows from different sources, support large-scale systems analyses in search of some adaptive solid management strategies, and assess not only technology-based options but also market-based instruments.

6. Conclusions

MSW management is normally seen as a major decision making issue with respect to sustainable development in all local communities of the EU. Due to the lack of appropriate system analysis methodologies to define, evaluate, optimize or adapt their waste treatment strategies and to meet the progressive targets set up at the EU level, this paper reviews all the possible trends, and evaluates the present situation of SWM systems in the EU countries in terms of waste processing systems, policy and decision making issues. Facing all regulatory agencies, industrial, and municipalities in the EU, whereas the Southern EU countries (e.g. Portugal, Greece, Spain) require developing measures to implement more integrative SWM systems and reach the objectives of the EU directives, the Central EU Countries (e.g. Germany, Austria, The Netherlands, United Kingdom, France) and certain Northern countries (e.g.

Norway) need models and tools in order to rationalize their technological choices and management strategies.

With a thorough literature review and elaborate investigation on how the system analysis techniques were developed and applied in these EU countries, a few future foci in research presented were also organized in this paper for public and private sectors to determine their future strength, thereby achieving the sustainability goals easily. These few milestones required for future development can be carved up front for EU members as follows:

- Deepen the structure of systems engineering models in the context of IMS which may incorporate more multi-faceted features covering economic, environmental, social, ecological, political; cultural, and managerial aspects for the sustainability assessment of current and future SWM systems.
- Provide synergistic tools to account for uncertainty associated with economic, environmental, social, ecological, political; cultural, and managerial aspects for SWM systems.
- Develop large-scale system analysis techniques in order to combine system assessment tools such as EIA, LCA, MFA, and even green accounting with systems engineering models such as optimization models to assess global warming potential, energy saving, and resources conservation practices so as to achieve sustainable waste management goals.
- Investigate both carbon and water footprints for all waste management alternatives as an integral part of system analysis leading to support complicated decision making and policy analyses under the global change impacts.
- Improve current waste management informatics such as MIS, DSS, and ES for the fulfillment of targets of environmental management and data reporting requirements to the EU in the context of cyber infrastructure applications.
- Conduct more cohesive CBA to support resources conservation plans such as Green Dot or the deposit-refund systems – packaging waste in SWM systems proposed by European Directives and incorporate more advanced assessments in terms of economic, environmental or social behavior of such systems with the aid of some system assessment tools such as LCA and MFA.
- Expand all ideas described above at different spatial and temporal scales.

With these efforts above, it is believed that the trends of current SWM systems and the perspectives of future SWM in association with the potential applications via integrating a plethora of different systems engineering models with a variety of system assessment tools should lead to improved insights and generate a suite of better management policies and strategies needed for the future.

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